



Application of environmental DNA metabarcoding in a lake with extensive algal blooms

Wu, Qianqian
Sakata, Masayuki
Wu, Deyi
Yamanaka, Hiroki
Minamoto, Toshifumi

(Citation)

Limnology, 22(3):363-370

(Issue Date)

2021-08

(Resource Type)

journal article

(Version)

Accepted Manuscript

(Rights)

This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's AM terms of use, but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at:...

(URL)

<https://hdl.handle.net/20.500.14094/0100482038>



Application of environmental DNA metabarcoding in a lake with extensive algal blooms

Authors: Qianqian Wu^{1*}, Masayuki K. Sakata^{1*}, Deyi Wu², Hiroki Yamanaka³,
Toshifumi Minamoto¹

¹ Graduate School of Human Development and Environment, Kobe University: 3-11
Tsurukabuto, Nada-ku, Kobe, Hyogo 657-8501, Japan

² School of Environmental Science and Engineering, Shanghai Jiao Tong University:
800 Dongchuan Rd., Shanghai 200-240, China

³ Faculty of Advanced Science and Technology, Ryukoku University: 1-5 Yokotani,
Seta-Oe Cho, Otsu, Shiga 520-2194, Japan

* Both authors equally contributed to the study.

Corresponding author: Qianqian Wu

Graduate School of Human Development and Environment, Kobe University: 3-11,
Tsurukabuto, Nada-ku, Kobe, Hyogo 657-8501, Japan

E-mail: wuqian1124@yahoo.co.jp

17 Tel: +81-78-803-7991

18 **Abstract**

19 Recently, environmental DNA (eDNA) metabarcoding techniques have been applied to
20 biodiversity investigations in aquatic ecosystems. However, no study has yet tested
21 whether this technique is effective for water bodies in which extensive algal blooms
22 break out. In this study, fish DNA metabarcoding was carried out in Lake Taihu, which
23 experiences extensive algal blooms, to confirm whether the technique is also effective
24 for fish diversity research in ecosystems with frequent and extensive blooms. In
25 December 2016, three samples were collected, including one collected in the presence
26 of algal blooms and two collected in the absence of algal blooms. In August 2017, six
27 samples were collected, including three collected in the presence of algal blooms and
28 three in the absence of algal blooms. Equal amount of water samples (1 L) were
29 collected from each site; however, the actual amount of filtrate varied with the site.
30 Twenty-seven freshwater fish species were detected from the water samples collected in
31 Lake Taihu. The results showed that the composition of the detected species did not
32 differ whether or not blooms were present. However, the amount of filtration could
33 influence the number of species detected. The results suggest that future eDNA

- 34 metabarcoding studies under similar water environments should increase the amount of
- 35 filtration to maximize number of species detected.
- 36 **Keywords:** algal blooms, environmental metabarcoding analysis, fish species
- 37 composition, Lake Taihu

38 **Introduction**

39 Global biodiversity is decreasing at an unprecedented rate, far beyond people's
40 imagination (Dirzo and Raven 2003). Human activity is one of the main causes of
41 biodiversity loss, which directly threatens natural ecosystems (Butchart et al. 2010).
42 Freshwater vertebrate species populations are declining at twice the rate of decline of
43 land or ocean vertebrates (Grooten and Almond 2018). Among the 29,500 freshwater
44 dependent species so far assessed in the IUCN Red List, 27% are currently at risk of
45 extinction (Tickner et al. 2020). Among aquatic organisms, fish species play an
46 important role in ecosystem linkages as well as commercial fisheries, yet many fish
47 species are in danger of extinction. In fact, in China, approximately 15% of freshwater
48 fish species are listed as endangered or vulnerable (199 out of the 1323 species
49 registered in freshwater systems) (Xing et al. 2016). Therefore, accurate assessment of
50 fish communities is key to maintaining freshwater fish diversity and freshwater
51 ecosystem protection, and it provides essential information for biological protection
52 (Kwak and Peterson 2007).

In recent years, environmental DNA (eDNA) analysis and its usage have received widespread attention. Environmental DNA represents all types of DNA found in the environment, including DNA from organisms in the form of metabolic waste, damaged tissue, or sloughed skin cells (Kelly et al. 2014; Barnes and Turner 2016). A method that retrieves DNA from environmental samples has been used to explore aquatic organisms in conservation and ecological studies (Pompanon et al. 2012; Bohmann et al. 2014; Thomsen and Willerslev 2015; Valentini et al. 20016). Two different approaches can be applied. One approach is eDNA analysis, which can detect a single species by using species-specific primers (Sean et al. 2019; Sakata et al. 2017; Wu et al. 2018, 2019; Fujiwara et al. 2016), while the other approach promotes diversity assessments of organisms using universal primers, called eDNA metabarcoding (Miya et al. 2015; Komai et al. 2019). The experimental procedure for eDNA metabarcoding includes collecting water samples and PCR amplification of the eDNA from fish species using universal primers (Miya et al. 2015; Deiner et al. 2016). The species composition was estimated by sequencing the PCR amplicons through next-generation sequencing (NGS). This technique is rapid and non-invasive method that is also cost-effective

69 (Deiner et al. 2017; Yamamoto et al. 2017). In freshwater fish species, eDNA
70 metabarcoding reveals comparable or higher number of fish species, compared to that
71 from traditional surveys (Hänfling et al. 2016; Shaw et al. 2016; Nakagawa et al. 2018).
72 In recent years, eDNA metabarcoding is used increasingly for characterizing the species
73 composition of ecological communities (Evans et al. 2016; Sato et al. 2017). Therefore,
74 we believe it would be suitable for studying fish biodiversity and composition in lakes
75 with serious eutrophication.

76 Eutrophication in reservoirs, estuaries, rivers, and lakes is widespread all over the
77 world, and the severity is increasing, especially in developing countries such as China
78 (Yang et al. 2008). When the nitrogen and phosphorus concentrations in a water body
79 reach certain levels, algal blooms can easily occur (Carpenter 2008; Vadeboncoeur,
80 Lodge and Carpenter 2001). Algal blooms can clog the gills of aquatic organisms
81 (Smith and Schindler 2009), and aquatic organisms such as fish and bottom-dwelling
82 animals cannot survive in oxygen-deprived aquatic environments as those in algal
83 blooms (Ansari, Gill and Khan 2010). Researchers studying such aquatic environments
84 have largely performed biodiversity surveys using traditional methods, and the

application of eDNA technology has emerged in recent years (Cai et al. 2017; Zhang et al. 2018). In the wake of increased environmental degradation and biodiversity loss, more biodiversity surveys will be required in eutrophic aquatic ecosystems with algal blooms to assess aquatic ecosystem health and the impact of the blooms on the ecosystems. Environmental DNA technology provides information on the presence or absence of organisms based on the genetic information obtained from environmental water samples. Most studies that apply metabarcoding techniques for eDNA analysis have been conducted in oligotrophic or mesotrophic waters. To date, only Handley et al. (2019) has applied the metabarcoding technique to investigate fish composition in eutrophic water. It has been reported that species that were sensitive to the trophic conditions of their habitat were distributed in areas with corresponding trophic conditions, while those that were not were widely distributed regardless of the trophic conditions (Handley et al. 2019). Eutrophic aquatic environments are vulnerable to massive algal blooms, and algae clog filter papers, reducing filtration capacity and organism detection rates in eDNA surveys. Nevertheless, very few studies have been conducted to evaluate the potential application of eDNA technology in aquatic

environments with extensive algal blooms on the surfaces. The present study was conducted in Lake Taihu, a representative highly eutrophic lake with extensive algal blooms, and the results of the study could reveal the feasibility of the survey approach in such environments.

This study aimed to determine whether the eDNA metabarcoding analysis method could be applied to a highly eutrophicated lake by analyzing the effect of algal blooms. As a test area, we selected Lake Taihu, China, a highly eutrophicated lake with frequent extensive algal blooms, where fish are an important aquatic resource. We compared the number of detected species and the differences in species composition with the presence or absence of algal blooms. Finally, species detection using eDNA metabarcoding was compared with the traditional methods and the use of the illustrated book for the fish of Lake Taihu.

Materials and Methods

Study site

The study area of Lake Taihu is located in the Yangtze River Delta in eastern China, with a surface area of 2,338 km² and an average depth of 1.89 m (Sun and Huang 1993). In the 1980s, there were between 72 and 106 different species of fish in the lake (Zhu et al. 2007). During the 2000s, massive blue-green algal blooms broke out in Lake Taihu (Wang and Shi 2008; Zhu 2008). These led to massive mortalities of fish and shrimp and triggered serious drinking water crises (Huang et al. 2014). In addition, due to the multiple impacts of overfishing and degradation of aquatic vegetation, significant changes took place in the fish community structure of the lake, so that now there are only 40 to 50 fish species in Lake Taihu (Zhu 2004).

Water samples were collected during summer and winter in the north of Lake Taihu. We surveyed three sites on December 21, 2016 (Fig. 1, blue circle), and six sites on August 27–28, 2017 (Fig. 1, red circle). The different sampling sites were selected as there were relatively few sites in winter with algal blooms present. Consequently, sampling was carried out across four sites with algal blooms and five sites with no algal blooms. We observed that algal blooms appeared in three sampling sites in summer and one site in winter (Table 1). At site T7, we conducted a survey in both summer (referred

as T7-S) and winter (T7-W). Two bottles of 500 mL water were collected at each site, and 500 µL of 10% (mass/volume) benzalkonium chloride solution was added to each sample to prevent DNA degradation (Yamanaka et al. 2017). We transported the water samples at ambient temperature and filtered them in the laboratory at Shanghai Jiao Tong University.

eDNA filtration and extraction

The water samples were filtered using 47-mm glass-fiber filters (GF/F; GE Healthcare Japan, Tokyo, Japan; nominal pore size = 0.7 µm; cf. Minamoto et al. (2016) for selection of filter type for eDNA sampling). In the present study, we unified the filtration labor, and filtration would continue until it could not be further carried out. The amount of sample filtration varied for each water sample because of filter clogging (Table 1). After filtering the samples, we filtered 1 L of ultrapure water to check for unintended cross-contamination in the laboratory (i.e., filtration blank). To prevent contamination, the equipment was cleaned with a bleach solution (diluted household bleach containing 0.1% sodium hypochlorite) for >5 min before reuse to remove

residual DNA. After filtration, the filters were stored at – 20 °C until DNA extraction.

Total eDNA was extracted from the filter using a DNeasy Blood and Tissue Kit

(Qiagen, Hilden, Germany) following Minamoto et al. (2019). After DNA extraction,

DNA was eluted from the DNeasy spin column with 100 µL of Buffer AE (Qiagen) and

stored at – 20 °C until metabarcoding analysis.

Amplicon library preparation and MiSeq sequencing

Amplicon libraries for fish eDNA metabarcoding were prepared using MiFish-U

primers (Miya et al. 2015). The first-round polymerase chain reaction (1st PCR) was

performed in a total reaction volume of 12 µL, containing 6.0 µL 2 × KAPA HiFi

HotStart ReadyMix (KAPA Biosystems, Wilmington, MA, USA), 3.6 pmol each of

MiFish-U primers, 1 µL eDNA template, and ultrapure water. The thermal cycle profile

was 95 °C for 3 min; 40 cycles of 98 °C for 20 s, 65 °C for 15 s, and 72 °C for 15 s; and

72 °C for 5 min. A polymerase chain reaction was performed with four repetitions for

each DNA sample, to reduce potential bias arising through a stochastic variation during

the PCR step. Ultrapure water was used instead of eDNA in four reaction mixtures

(non-template negative controls). After four technical replicates of 1st PCR products were pooled into a single tube, unreacted reagents and primer dimers were removed from the 1st PCR products with the SPRI select Reagent Kit (Beckman Coulter, Brea, CA, USA) according to the manufacturer's instructions. The purified 1st PCR product concentration of each sample was measured by Qubit fluorometer 3.0 (Thermo Fisher Scientific, Waltham, MA, USA) using Qubit dsDNA HS assay kit. Then, all samples were diluted to 0.1 ng/μL, and all negative controls were diluted using an average dilution ratio.

The second-round PCR (2nd PCR) was performed to add MiSeq adapter sequences and 8-bp index sequences to both amplicon ends (Hamady et al. 2008). The total reaction volume of the 2nd PCR was also 12 μL, containing 6.0 μL of 2 × KAPA HiFi HotStart ReadyMix (KAPA Biosystems), 3.6 pmol each of forward and reverse primers, 1 μL template, and ultrapure water. The thermal cycle profile for the 2nd PCR was 95 °C for 3 min; 12 cycles of 98 °C for 20 s, 72 °C for 20 s, and 72 °C for 5 min. The 2nd PCR products were pooled in one tube (i.e., one pooled 2nd PCR products that included all samples). The library sample of target size amplicons was obtained by

electrophoresis using E-Gel SizeSelect 2% (ThermoFisher Scientific) with the E-Gel Precast Agarose Electrophoresis System (ThermoFisher Scientific). Subsequently, it was confirmed whether only DNA of the target length (approximately 370 bp) was isolated by using an Agilent 2100 Bioanalyzer (Agilent Technologies, Santa Clara, CA, USA). The concentration of the DNA library was adjusted to 4 nM. Finally, the library was sequenced using an Illumina MiSeq v2 Reagent kit for 2 × 150 bp PE (Illumina, San Diego, CA, USA).

Sequence data analysis

The raw reads obtained from MiSeq sequencing were preprocessed and analyzed using USEARCH v10.0.240 (Edgar 2010). First, R1 and R2 reads were assembled using the command “-fastq_mergepairs”. Second, the primer sequence was removed from both edges of the assembled reads by using commands “-fastx_truncate”, “-stripelleft” and “-stripelright”. Third, low quality reads with an expected error rate (Edgar and Flyvbjerg 2015) of >1% and too short reads of <140 bp were removed by quality filtering (using command “-fastq_filter”). Fourth, the preprocessed reads were dereplicated using the

196 “fastx_uniques” command. Fifth, the dereplicated reads were denoised using the
197 “unoise3” command to generate amplicon sequence variants (ASVs) that remove all
198 putatively chimeric, erroneous sequences (Edgar 2016) and partial ASVs with fewer
199 than 10 reads. Finally, ASVs were subjected to taxonomic assignments to species names
200 using the “BLAST” search program to nucleotide database using a nucleotide query
201 ([https://blast.ncbi.nlm.nih.gov/Blast.cgi?PROGRAM=blastn&PAGE_TYPE=BlastSearch](https://blast.ncbi.nlm.nih.gov/Blast.cgi?PROGRAM=blastn&PAGE_TYPE=BlastSearch&LINK_LOC=blasthome)
202 [ch&LINK_LOC=blasthome](https://blast.ncbi.nlm.nih.gov/Blast.cgi?PROGRAM=blastn&PAGE_TYPE=BlastSearch&LINK_LOC=blasthome)). In the next step, the search was performed with default
203 settings, and ASVs of sequence identity >98.5% (two nucleotide differences allowed)
204 were judged as that species or was judged a species of that genus in the case of no
205 distinction at the species level. When ASVs of sequence identity of <98.5% were
206 assigned to top hit species, and when the sequence identity was >90% to reference
207 sequences, ASVs were classified as belonging to the genus or family level. In addition,
208 the detected reads in the negative control had possibly been contaminated, and the
209 number of *Carassius spp.* reads detected in the filtration blank of 2016 was removed
210 from reads of all corresponding environmental samples from 2016 (Port et al. 2016;
211 Hayami et al. 2020).

212

213 **Statistical analysis**

214 All analyses were performed using R ver. 3.6.0 (R Development Core Team 2019), the
215 vegan package version 2.5-6 (Oksanen et al. 2019), and the lme4 package version 1.1-
216 21 (Bates et al. 2015). The result of the species accumulation curve generated by
217 “rarecurve” function in the vegan package shows that the number of detected species in
218 each sample is saturated, except for the sample T2. In all analyses, the reads data were
219 converted to presence/absence (Jeunen et al. 2019; Fig. S1).

220 A rarefaction analysis was performed to account for the variations in the number
221 of total reads across samples (Fig S2). The following two analyses were performed
222 individually for un-rarefied and rarefied data, which was obtained by unifying the
223 number of total reads to the fewest reads through rarefying, using the R package
224 “vegan” version 2.5-6. To estimate whether algal blooms (presence/absence) affect the
225 number of fish species detected by eDNA metabarcoding, we used a generalized linear
226 mixed model (GLMM) with a poisson distribution. In the model, we assigned the
227 absence of algal blooms = 0 and the presence of algal blooms = 1 and set the

228 presence/absence of algal blooms as explanatory variables, the number of fish species
229 detected by eDNA metabarcoding as a response variable, filtration volume as an offset
230 term, and site IDs as random effects. Regarding the sampling sites in the present
231 analysis, some studies have reported that the fish community in Lake Taihu is evenly
232 distributed throughout the entire lake since being affected by severe environmental
233 change and overfishing (Tao et al. 2010; Mao et al. 2011). Therefore, we assumed that
234 there was no bias in potential fish communities based on sampling location or season,
235 and the site IDs was randomized in the analysis. Non-metric multidimensional scaling
236 (NMDS) was then performed with “Jaccard methods” and 10,000 permutations to
237 visualize the dissimilarity of fish composition by the presence/absence of algal blooms.
238 In addition, to compare fish composition between the presence and absence of algal
239 blooms, we performed PERMANOVA by function “adonis” in the vegan package. In
240 this analysis, all the data were used to conduct community analysis, because similar
241 result was obtained even at site T2, where only one ASV was detected, was removed
242 (Fig. S3). Two mammals (*Homo sapiens* and *Sus scrofa*) were detected in the eDNA

metabarcoding analysis. Mammals were excluded from the statistical analysis. Parts of graphs were drawn using the ggplot2 package version 3.1.1 (Wickham 2019).

Finally, the fish composition obtained from eDNA and the latest traditional survey were compared (eDNA survey: this study, traditional survey and the illustrated handbook for the fish of Lake Taihu: Mao et al. 2011; Li and Shimatani 2016).

Results

In total, 1,207,279 MiSeq reads were obtained. After bioinformatic filtering (see Materials and Methods), 465,347 reads were retained, corresponding to 38.55% of the total reads. From the filtration blank in 2016, 55 reads of *Carassius spp.* were detected, while no fish DNA was detected in the filtration blank in 2017. In total, 28 freshwater fish species were detected (Table S1).

The results for the un-rarefied data showed that the number of detected fish species was significantly lower in the presence of algal blooms (GLMM, $p = 0.016$; Fig. 2). In addition, there was no significant difference in the composition of detected species between the presence/absence of algal blooms ($p = 0.969$; Fig. 3). The results

for rarefied data indicated the significantly lower number of detected species in the presence of algal blooms (GLMM, $p=0.014$; Fig. S4) and non-significant difference in the composition of species ($p=0.514$; Fig. S5). Therefore, rarefaction did not alter the results. Additionally, among the 27 freshwater fish species detected by eDNA metabarcoding analysis, 10 species appeared in the traditional survey, and 13 species appeared in the illustrated handbook for the fish of Lake Taihu (Li and Shimatani 2016) (Fig. 4). According to the species accumulation curve, the number of detected species did not reach a saturation by metabarcoding, after sampling nine sites (Fig. S6).

Discussion

This study attempted to assess the fish diversity in a severely eutrophicated lake with extensive algal blooms using eDNA metabarcoding. Our results showed that fish species could be detected even in eutrophicated areas where extensive algal blooms break out; however, in such an environment, besides the existence of algae blooms, other factors (e.g., filtration volume and PCR inhibitors) could affect the number of

species detected. In future research, how the aforementioned factors influence the number of fish species detected should be explored.

Although the composition of detected species did not differ by the presence or absence of algal blooms, the number of detected species showed significant differences (Fig. 2, 3). There could be two reasons why the composition of the detected species was not affected. First, since the late 1980s, serious eutrophication has been occurring in Taihu Lake, and fish species sensitive to oxygen-poor environments have been significantly reduced, while those adapted to oxygen-poor environments have survived (Zhu, 2004; Mao et al. 2011). It has also been reported that some fish, such as *Coilia* and *Carassius*, can survive by changing their diet (e.g., consume cyanobacteria) (Liu, 2009). As a result, only fish species with strong adaptability to oxygen-poor conditions may remain in such an unfavorable environment. Therefore, the presence or absence of algal blooms may not influence species composition in the lake. In the present study, the filtration volume of the water samples at sites without algal blooms was significantly higher than at sites with algal blooms (Table 1; Fig. S7). When investigating fish diversity in water bodies with extensive and frequent algal blooms

using eDNA, algae could clog filters, and, in turn, directly influence filtration yield and indirectly influence DNA recovery. Consequently, the number of species detected could be relatively low, and saturation points may not be achieved in study sites. The obtained reads, classified as fish sequences after filtering, were compared; lower reads were obtained for the sites with algal blooms (Table S1). This might be attributed to the limitations in the PCR amplification, such as PCR inhibition in the samples from sites with algal blooms. Although previous studies have shown that sample dilution could circumvent the effects of inhibition (Biggs et al. 2015; Cilleros et al. 2019), no dilution was performed in the present study because the amount of water that could be filtered was very small and fish DNA concentrations in the samples were assumed to be very low. Previous studies have reported positive relationships between the number of detected species and filtered water volume (Mächler et al. 2016; Sakata et al. 2020). Consequently, increasing the amount of filtration is key for improving species detection rates. This is the first investigation in the water environment rich in algae, and the filtration has been performed according to the standard experimental methods stipulated before. In future studies, multiple filter papers, filter with a larger pore size, or pre-

306 filtration to treat water containing more particles are required. Alternatively, the
307 subsurface water under algal blooms could also be collected as opposed to the surface
308 water covered by algal blooms; this could improve the efficiency of eDNA
309 metabarcoding in eutrophic environments. Meanwhile, some studies have demonstrated
310 DNA in sediment has higher concentrations and can be conserved for longer periods
311 than DNA in water (Corinaldesi et al. 2011; Sakata et al. 2021). Therefore, collecting
312 both water samples and sediment samples could enhance detection capacity.

313 MiFish marker is generally useful for identifying fish species in Lake Taihu. A
314 total of 104 fish species were found in Lake Taihu using traditional methods and an
315 illustrated fish handbook in Taihu Lake (Mao et al. 2011; Li and Shimatani 2016).
316 Excluding 17 species that had no reference data, the rest of the 87 species had reference
317 sequences; however, some species were indistinguishable because the sequences in the
318 marker regions were identical among species, so that some species were not easily
319 identifiable in our data (Table S1). Some DNA sequences were not successfully
320 identified because of high similarity within the genus, resulting in the genus-level
321 classification (i.e., *Carassius*, *Hypophthalmichthys*). The similarity of DNA sequences

of cyprinid fish is high within and between genus, which leads to the subfamily level classification (i.e., *Acanthohodeus* and *Cultrinae*). We cannot rule out the possibility that the deposited sequence on NCBI may be wrong or caused by incorrect species identification. Therefore, in future research, it will be essential to establish a perfectly correct database. At the same time, we also found that the sequences of many fishes living in Lake Taihu are very close and cannot be fully identified by using 12S rRNA fragments. It is also suggested that a longer marker for metabarcoding or specific primers for a certain genus is required to solve this identification problem.

Environmental DNA metabarcoding analysis has been applied to biomonitoring and has been used to investigate the effect of external factors such as water quality. Our results indicate that the presence of algal blooms did not affect the community structure detected by eDNA metabarcoding, but the number of fish species. A body of essential information of this nature will assist in designing eDNA metabarcoding biomonitoring programs and their application in complex environments, contributing to aquatic biodiversity conservation.

338 **Acknowledgments**

339 We are grateful to the students of Kobe University and Shanghai Jiao Tong University,
340 who helped with water sampling. This work was partly supported by the Japan Society
341 for Promotion of Science (JSPS) KAKENHI (Grant numbers: JP17H03735 and
342 JP20H03326) and by the Environmental Research and Technology Development Fund
343 (JPMEERF20S20704) of the Ministry of the Environment, Japan.

344 **References**

- 345 1. Ansari AA, Gill SS, Khan FA (2010) Eutrophication: Threat to Aquatic
346 Ecosystems. Pages 143-170 *in* Eutrophication: causes, consequences and control.
347 Springer, Dordrecht.
- 348 2. Barnes AM, Turner CR (2016) The ecology of environmental DNA and implications
349 for conservation genetics. *Conserv Genet* 17:1–17.
- 350 3. Bates D, Mächler M, Bolker BM, Walker SC (2015) Fitting linear mixed-effects
351 models using lme4. *J Stat Softw* 67:1–48.
- 352 4. Biggs J, Ewald N, Valentini A, Gaboriaud C, Dejean T, Griffiths RA, Foster J,
353 Wilkinson JW, Arnell A, Brotherton P, Williams P, Dunn F (2015) Using eDNA to
354 develop a national citizen science-based monitoring programme for the great crested
355 newt (*Triturus cristatus*). *Biol Conserv* 183:19–28.
- 356 5. Bohmann K, Evans A, Gilbert MTP, Carvalho GR, Creer S, Knapp M, Yu DW,
357 Bruyn M de (2014) Environmental DNA for wildlife biology and biodiversity
358 monitoring. *Trends Ecol Evol* 29:358–367.
- 359 6. Butchart SHM, Walpole M, Collen B et al (2010) Global biodiversity: indicators of
360 recent declines. *Sci* 328:1164–1168.
- 361 7. Carpenter SR (2008) Phosphorus control is critical to mitigating eutrophication.
362 *PNAS*, 105:11039–11040.
- 363 8. Ciller K, Valentini A, Allard L, Dejean T, Etienne R, Grenouillet G, Iribar A,
364 Taberlet P, Vigouroux R, Brosse S (2019) Unlocking biodiversity and conservation
365 studies in high-diversity environments using environmental DNA (eDNA): A test
366 with Guianese freshwater fishes. *Mol Ecol Resour*, 19:27–46.
- 367 9. Dirzo R, Raven PH (2003) Global state of biodiversity and loss. *Annu Rev Environ*
368 *Resour*, 28:137–167.
- 369 10. Deiner K, Bik HM, Mächler E, Seymour M, Lacoursière-Roussel A, Altermatt F,
370 Creer S, Bista I, Lodge DM, Vere N de, Pfrender ME, Bernatchez L (2017)

- 371 Environmental DNA metabarcoding: Transforming how we survey animal and
372 plant communities. *Mol Ecol* 26:5872–5895.
- 373 11. Deiner K, Fronhofer EA, Mächler E, Walser JC, Altermatt F (2016) Environmental
374 DNA reveals that rivers are conveyor belts of biodiversity information. *Nat Commun*
375 7:12544.
- 376 12. Dudgeon D, Arthington AH, Gessner MO, Kawabata Z, Knowler DJ, Lévêque C,
377 Naiman RJ, Prieur-Richard AH, Soto D, Stiassny MLJ, Sullivan CA (2006)
378 Freshwater biodiversity: importance, threats, status and conservation challenges.
379 *Biol Rev* 81:163–182.
- 380 13. Edgar RC (2010) Search and clustering orders of magnitude faster than BLAST.
381 *Bioinformatics*. 26:2460–2461.
- 382 14. Edgar RC, Flyvbjerg H (2015) Error filtering, pair assembly and error correction
383 for next-generation sequencing reads. *Bioinformatics* 31:3476–3482.
- 384 15. Edgar RC (2016) UNOISE2: improved error-correction for Illumina 16S and ITS
385 amplicon sequencing. *bioRxiv*:081257.
- 386 16. Evans NT, Olds BP, Renshaw MA, Turner CR, Li Y, Jerde CL, Mahon AR,
387 Pfrender ME, Lamberti GA, Lodge DM (2016) Quantification of mesocosm fish
388 and amphibian species diversity via environmental DNA metabarcoding. *Mol Ecol*
389 *Resour* 16:29–41.
- 390 17. Fujiwara A, Matsushashi S, Doi H, Yamamoto S, Minamoto T (2016) Use of
391 environmental DNA to survey the distribution of an invasive submerged plant in
392 ponds. *Freshw Sci* 35:748–754.
- 393 18. Gardner R, Almond R (2018) Living Planet Report 2018: Aiming Higher. World
394 Wildlife Fund.
- 395 19. Hamady M, Walker JJ, Harris JK, Gold NJ, Knight R (2008) Error-correcting
396 barcoded primers for pyrosequencing hundreds of samples in multiplex. *Nat*
397 *Methods* 5:235–237.
- 398 20. Hänfling B, Lawson Handley L, Read DS, Hahn C, Li J, Nichols P, Blackman RC,

- Oliver A, Winfield IJ (2016) Environmental DNA metabarcoding of lake fish communities reflects long-term data from established survey methods. *Mol Ecol* 25:3101–3119.
21. Handley LL, Read DS, Winfield IJ, Kimbell H, Johnson H, Li J, Hahn C, Blackman R, Wilcox R, Donnelly R, Szitenberg A, Hänfling B (2019) Temporal and spatial variation in distribution of fish environmental DNA in England’s largest lake. *Environmental DNA* 1:26–39.
22. Hayami K, Sakata MK, Inagawa T, Okitsu J, Katano I, Doi H, Nakai K, Ichiyanagi H, Gotoh RO, Miya M, Sato H, Yamanaka H, Minamoto T (2020) Effects of sampling seasons and locations on fish environmental DNA metabarcoding in dam reservoirs. *Ecol & Evol* 10:5354–5367.
23. Huang C, Li Y, Yang H, Sun D, Yu Z, Zhang Z, Chen X, Xu L (2014) Detection of algal bloom and factors influencing its formation in Taihu Lake from 2000 to 2011 by MODIS. *Environ Earth Sci* 71:3705–3714.
24. Jeunen G, Lamare M D, Knapp M, Spencer HG, Taylor HR, Stat M, Bunce M, Gemmell NJ (2019) Water stratification in the marine biome restricts vertical environmental DNA (eDNA) signal dispersal. *Environmental DNA* 2:99–111.
25. Kelly RP, Port JA, Yamahara KM, Crowder LB (2014) Using Environmental DNA to Census Marine Fishes in a Large Mesocosm. *PLoS One* 9, e86175.
26. Komai T, Gotoh RO, Sado T, Miya M (2019) Development of a new set of PCR primers for eDNA metabarcoding decapod crustaceans. *MBMG* 3:1–19.
27. Kwak TJ, Peterson JT (2007) Community indices, parameters, and comparisons. Pages 677–763 *in* Guy CS and Brown ML (editor). *Analysis and Interpretation of Freshwater Fisheries Data*. American Fisheries Society, Bethesda, MD.
28. Li J, Shimatani Y (2016) *Illustrated fish handbook in the east Tiaoxi river*. Science press, China (in Chinese).
29. Liu E, Liu Z, Chen W, Chen K (2005) Changes in the yield and composition of the fish catches and their relation to the environmental factors in Lake Taihu. *Lake Sci* 17:251–255(in Chinese).

30. Liu E (2009) Changes of fish community, the mechanism of changes, and the effects of the changes on environment in Taihu Lake. *Journal of Hydroecology* 30:8–14 (in Chinese).
31. Mächler E, Deiner K, Spahn F, Altermatt F (2016) Fishing in the water: effect of sampled water volume on environmental DNA-based detection of macroinvertebrates. *Environ Sci Technol* 50:305–312.
32. Mao Z, Hu X, Zeng Q, Zhou L, Wang X, Wu L, Cao P, Sun M (2011) Community structure and diversity of fish in Lake Taihu. *Chinese J Ecol* 30:2836–2842.
33. Minamoto T, Naka T, Moji K, Maruyama A (2016) Techniques for the practical collection of environmental DNA: filter selection, preservation, and extraction. *Limnology* 17:23–32.
34. Minamoto T, Hayami K, Sakata MK, Imamura A (2019) Real-time PCR assays for environmental DNA detection of three salmonid fish in Hokkaido, Japan: application to winter surveys. *Ecol Res* 34:237–242.
35. Miya M, Sato Y, Fukunaga T, Sado T, Poulsen JY, Sato K, Minamoto T, Yamamoto S, Yamanaka H, Araki H, Kondoh M, Iwasaki W (2015) MiFish, a set of universal PCR primers for metabarcoding environmental DNA from fishes: detection of more than 230 subtropical marine species. *R Soc Open Sci* 2:150088.
36. Nakagawa H, Yamamoto S, Sato Y, Sado T, Minamoto T, Miya M (2018) Comparing local- and regional-scale estimations of the diversity of stream fish using eDNA metabarcoding and conventional observation methods. *Freshw Biol* 63:569–580.
37. Oksanen J, Blanchet FG, Friendly M, Kindt R, Legendre P, McGlinn D, ... Wagner H (2019) *vegan: community ecology package*. R package version 2.5-6.
38. Port JA, O'Donnell JL, Romero-Maraccini OC, Leary PR, Litvin SY, Nickols KJ, Yamahara KM, Kelly RP (2016) Assessing vertebrate biodiversity in a kelp forest ecosystem using environmental DNA. *Mol Ecol* 25:527–541.
39. Pompanon F, Deagle BE, Symondson WOC, Brown DS, Jarman SN, Taberlet P (2012) Who is eating what: diet assessment using next generation sequencing. *Mol*

- 457 Ecol 21:1931–1950.
- 458 40. R Core Team 2019. R: A language and environment for statistical computing. R
459 Foundation for Statistical Computing, Vienna, Austria.
- 460 41. Sard NM, Herbst SJ, Nathan L, Uhrig G, Kanefsky J, Robinson JD, Scribner KT
461 (2019) Comparison of fish detections, community diversity, and relative abundance
462 using environmental DNA metabarcoding and traditional gears. Environmental
463 DNA 1:368–384.
- 464 42. Sakata MK, Maki N, Sugiyama H, Minamoto T (2017) Identifying a breeding
465 habitat of a critically endangered fish, *Acheilognathus typus*, in a natural river in
466 Japan. Sci Nat 104:11–12.
- 467 43. Sakata MK, Yamamoto S, Gotoh RO, Miya M, Yamanaka H, Minamoto T (2020)
468 Sedimentary eDNA provides different information on timescale and fish species
469 composition compared with aqueous eDNA. Environmental DNA 2:505–518.
- 470 44. Sakata MK, Watanabe T, Maki N, Ikeda K, Kosuge T, Okada H, Yamanaka H, Sado
471 T, Miya M, Minamoto T (2021) Determining an effective sampling method for
472 eDNA metabarcoding: a case study for fish biodiversity monitoring in a small,
473 natural river. Limnology 21:221–235.
- 474 45. Sato H, Sogo Y, Doi H, Yamanaka H (2017) Usefulness and limitations of sample
475 pooling for environmental DNA metabarcoding of freshwater communities. Sci Rep
476 7:14860.
- 477 46. Sean NG, McKenzie MB, Drake J, Harper LR, Browett SS, Coscia I, Wangenstein
478 OS, Baillie C, Bryce E, Dawson DA, Ochu E, Hänfling B, Handley LL, Mariani S,
479 Lambin X, Sutherland C, McDevitt AD (2019) Fishing for mammals: landscape-
480 level monitoring of terrestrial and semi-aquatic communities using eDNA from
481 lotic ecosystems. bioRxiv:629758.
- 482 47. Shaw JL, Clarke LJ, Wedderburn SD, Barnes TC, Weyrich LS, Cooper A (2016)
483 Comparison of environmental DNA metabarcoding and conventional fish survey
484 methods in a river system. Biol Conserve 197:131–138.

- 485 48. Smith VH, Schindler DW (2009) Eutrophication science: where do we go from
486 here? *Trends Ecol Evol* 24:202–207.
- 487 49. Sun SC, Huang YP (1993) Lake Taihu. Ocean Press, Beijing, 273 (in Chinese).
- 488 50. Thomsen PF, Willerslev E (2015) Environmental DNA – An emerging tool in
489 conservation for monitoring past and present biodiversity. *Biol Conserv* 183:4–18.
- 490 51. Tickner D, Opperman JJ, Abell R, Acreman M, Arthington AH, Bunn SE, Cooke
491 SJ, Dalton J, Darwall W, Edwards G, Harrison I, Hughes K, Jones T, Leclère
492 D, Lynch AJ, Leonard P, McClain ME, Muruven D, Olden JD, Ormerod
493 SJ, Robinson J, Tharme RE, Thieme M, Tockner K, Wright M, Young L (2020)
494 Bending the Curve of Global Freshwater Biodiversity Loss: An Emergency
495 Recovery Plan. *BioScience* 70: 330–342.
- 496 52. Vadeboncoeur Y, Lodge DM, Carpenter SR (2001) Whole-lake fertilization effects
497 on distribution of primary production between benthic and pelagic habitats.
498 *Ecology* 82:1065–1077.
- 499 53. Valentini A, Taberler P, Miaud C, Civadee R, Herder J, Thomsen PF, Bellemain
500 E, Besnard A, Coissac E, Boyer F, Boyer F, Gaboriaud C, Jean P, Poulet N, Roset
501 N, Copp GH, Geniez P, Pont D, Argillier C, Baudoin J, Peroux T, Crivelli AJ,
502 Olivier A, Acqueberge M, Brun ML, Møller PR, Willerslev E, Dejean T (2016)
503 Next-generation monitoring of aquatic biodiversity using environmental DNA
504 metabarcoding. *Mol Ecol* 25:929–942.
- 505 54. Wang M, Shi W (2008) Satellite-observed algae bloom in China's Lake Taihu. *EOS*
506 *Trans AGU* 89:201–202.
- 507 55. Wickham H (2019) ggplot2: Create elegant data visualisations using the grammar of
508 graphics. R package version 3.1.1 [Online] Available at: [https://cran.r-](https://cran.r-project.org/web/packages/ggplot2/index.html)
509 [project.org/web/packages/ggplot2/index.html](https://cran.r-project.org/web/packages/ggplot2/index.html).
- 510 56. Wu Q, Kawano K, Ishikawa T, Sakata MK, Nakao R, Hiraiwa MK, Tsuji S,
511 Yamanaka H, Minamoto T (2019) Habitat selection and migration of the common

- 512 shrimp, *Palaemon paucidens* in Lake Biwa, Japan-An eDNA-based study.
 513 Environmental DNA 1:54–63.
- 514 57. Wu Q, Kawano K, Uehara Y, Okuda N, Hongo M, Tsuji S, Yamanaka H, Minamoto
 515 T (2018) Environmental DNA reveals nonmigratory individuals of *Palaemon*
 516 *paucidens* overwintering in Lake Biwa shallow waters. Freshw Sci 37:307–314.
- 517 58. Xing Y, Zhang C, Fan E, Zhao Y (2016) Freshwater fishes of China: species
 518 richness, endemism, threatened species, and conservation. Diversity Distrib 22:
 519 358–370.
- 520 59. Yamamoto S, Masuda R, Sato Y, Sado T, Araki H, Kondoh M, Minamoto T, Miya
 521 M (2017) Environmental DNA metabarcoding reveals local fish communities in a
 522 species-rich coastal sea. Sci Rep 7:40368.
- 523 60. Yamanaka H, Minamoto T, Matsuura J, Sakurai S, Tsuji S, Motozawa H, Hongo
 524 M, Sogo Y, Kakimi N, Teramura I, Sugita M, Baba M, Kondo A (2017) A simple
 525 method for preserving environmental DNA in water samples at ambient
 526 temperature by addition of cationic surfactant. Limnology 17:233–241.
- 527 61. Yang X, Wu X, Hao H, He Z (2008) Mechanisms and assessment of water
 528 eutrophication. Journal of Zhejiang University SCIENCE 9:197–208.
- 529 62. Zhu GW (2008) Eutrophic status and factors causing a large, shallow, and
 530 subtropical lake in Taihu, China. Lake Sci 20:21–26 (in Chinese).
- 531 63. Zhu SQ (2004) Ichthyological survey of Lake Taihu during 2002–2003. Lake
 532 Sci 16:120–124 (in Chinese).
- 533 64. Zhu SQ, Liu ZW, Gu XH (2007) Changes in fish fauna and fish yield analysis in
 534 Lake Taihu. Lake Sci 19:664–669 (in Chinese).

535 **Table 1.** Filtration volume of each water sample. + means that we observed the presence of algal
 536 blooms at the site and – means that we did not.

Season	Date	Site	Filtration volume	Algal blooms
Winter	Dec-21, 2016	T5	400ml	-
	Dec-21, 2016	T7-W	400ml	-
	Dec-21, 2016	T8	100mL	+
Summer	Aug-27, 2017	T1	350mL	+
	Aug-27, 2017	T2	155mL	+
	Aug-28, 2017	T3	1000mL	-
	Aug-28, 2017	T4	22.5mL	+
	Aug-28, 2017	T6	1000mL	-
	Aug-28, 2017	T7-S	560mL	-

537

Figure legends

Figure 1.

Map of the study sites. Green squares mean that we observed the presence of algal blooms at the site. Red and blue circles show the sampling sites in summer and winter, respectively. A red-edged white circle means no fish ASVs have been detected. Circle size indicates the number of detected species. With regard to T7, we conducted surveys in both summer (T7-S) and winter (T7-W).

Figure 2.

Number of fish species detected in the presence/absence of algal blooms by environmental DNA metabarcoding. The GLMM result shows that number of species is significant higher when algal blooms are absent than when they are present.

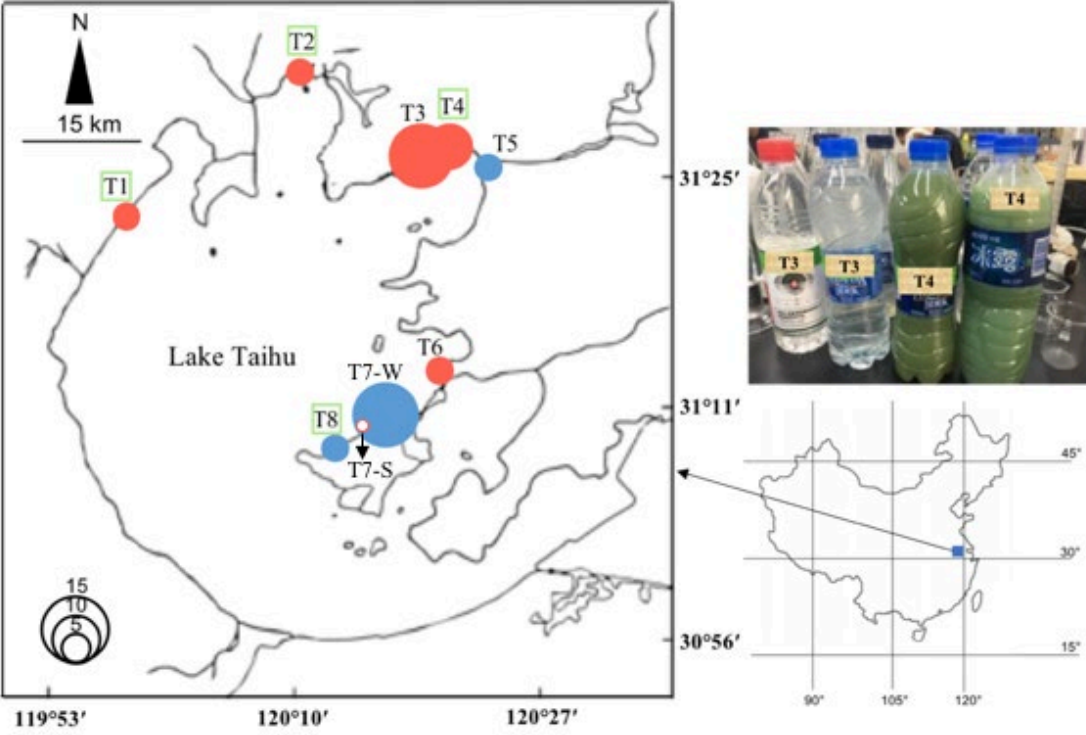
Figure 3.

Composition of detected fish species in the presence/absence of algal blooms. Circles and grey solid triangle represent the samples with algal blooms ($N = 4$), crosses and white dotted triangle represent the samples without algal blooms ($N = 5$). No fish species was detected from the sample of T7-S without algal blooms, and therefore, only four points for the latter were plotted.

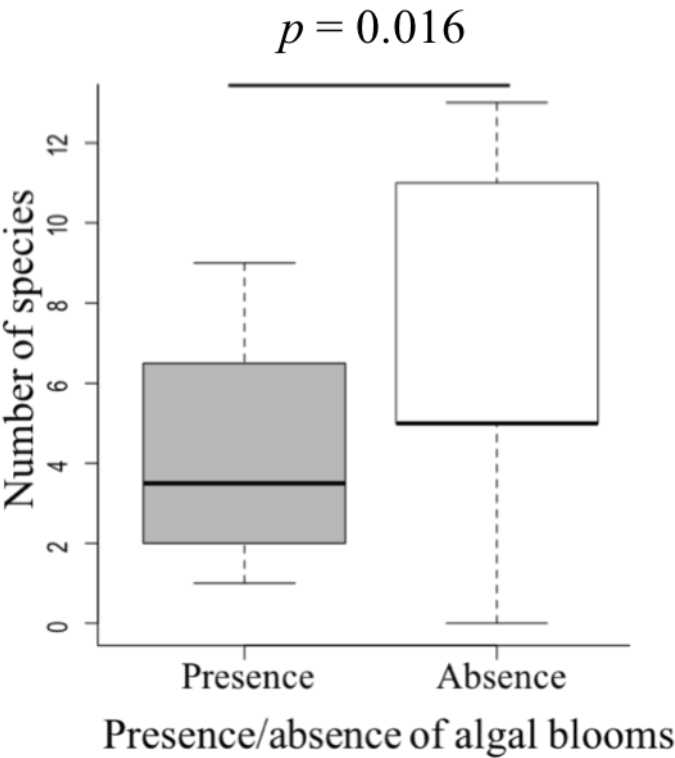
Figure 4.

Comparison of the results of detected fish in Lake Taihu confirmed by traditional methods and by the environmental DNA (eDNA) method. The upper-left circle represents the result from eDNA, upper-right represents the fish caught by trawl net (Mao et al. 2011), and the bottom shows the fish species in the illustrated handbook of Lake Taihu (Li and Shimatani 2016).

563 **Figures**
564 **Figure 1.**



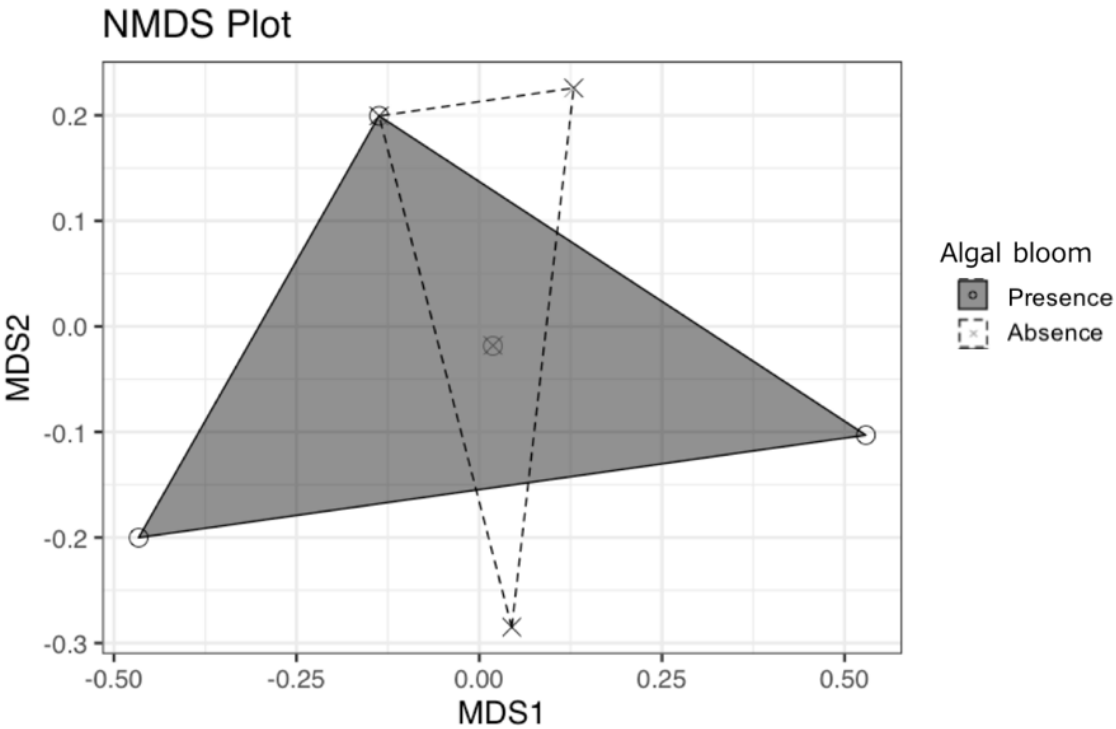
566 **Figure 2.**



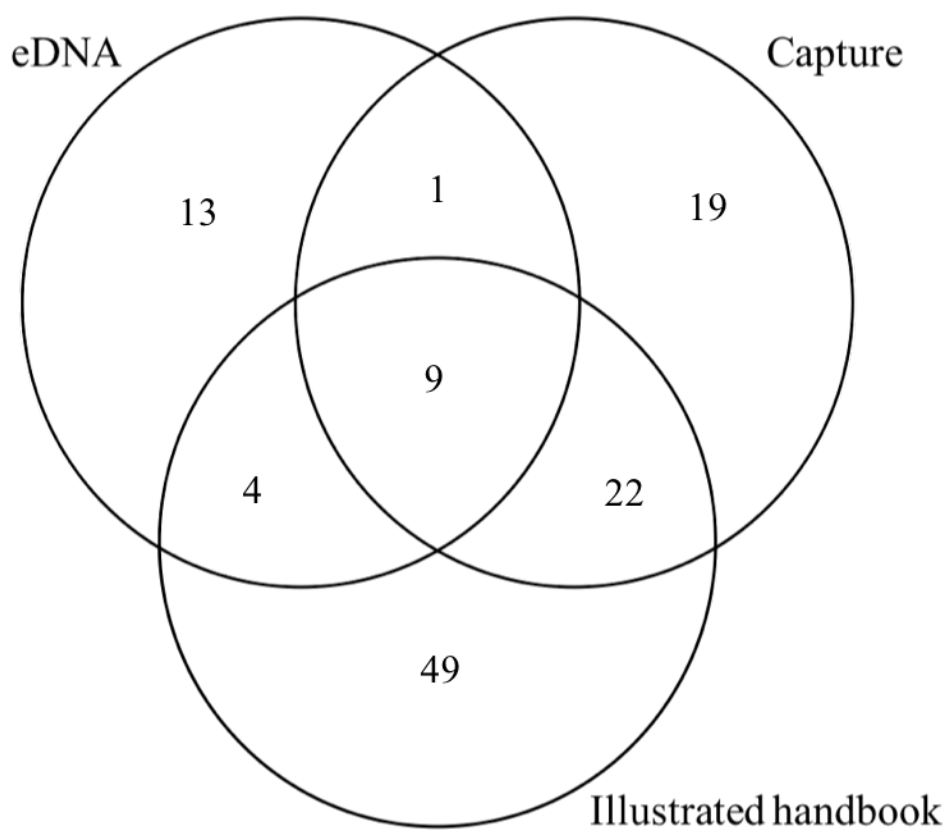
567

568

569 **Figure 3.**



571 **Figure 4.**



572